1 Temporal bisphenol A occurrence in the Danube surface water during different seasons

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11 Abstract:

This research describes for the first time the identification and quantification ofbisphenol A (BPA) in the Danube surface water in Novi Sad (Serbia).

A gas chromatographic mass spectrometric (GC-MS) method with a solid-phase extraction (SPE) procedure was used to investigate the seasonal variations of the BPA concentrations in the water samples. These variations were significant, showing higher concentrations in summer comparing with those in winter. Although it has a short half-life in surface water (0.5-5 days), BPA was frequently detected at relevant concentration levels of up to 693 ng/L.

The risk quotients, expressed as the ratio of the 95th percentile of the measured environmental concentrations for each sampling site and the predicted no-effect concentration (PNEC) were higher than one (> 1) at seven sampling locations, indicating a significant ecotoxicological risk to the human health and the environment, which confirmed a considerable contamination of the Danube surface water along the Novi Sad bank, Serbia.

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Keywords: bisphenol A; the Danube river; endocrine disrupter; GC-MS; surface water.

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27 **RUNNING TITLE:** BISPHENOL A IN THE DANUBE

INTRODUCTION

29

Bisphenol A (BPA) is one of the most investigated and discussed endocrine disrupting 30 compounds (EDCs), and has received large attention as an emerging chemical of concern in 31 order to protect wildlife and human health.¹ Endocrine disrupters (EDs) can interfere with 32 hormonal action and with the human cell functions mainly through the interaction with 33 specific receptors and can induce numerous adverse health effects. Therefore, the occurrence 34 of BPA in the environment could have potential harmful effects on the reproductive system, 35 growth, digestion and metabolic processes of aquatic organisms like fish or mussels.² 36 Moreover, effects on the human endocrine system and its function have been discussed by 37 direct exposure via drinking water (BPA was used in polycarbonate plastic bottles).³ BPA is 38 also believed to pose a potential risk of developing various types of cancers.^{4,5} 39

BPA, 2,2-bis(4-hydroxyphenyl) propane, is an organic compound extensively used as 40 an intermediate in the production of polycarbonate plastics and epoxy resins, unsaturated 41 polyester-styrene resins and flame retardants. It is used in a broad range of the products of 42 modern consumer society, including reusable baby bottles, toys, food and drink storage 43 containers, protective coatings for metal cans, adhesives and paints, electronic components, 44 45 personal computers, CDs and DVDs, sports equipment, medical devices, eyeglass lenses, dental sealants and dental bonding agents, etc.^{6,7} More than 3.6 million tons of BPA is 46 synthesized worldwide annually and its production is constantly increasing owing to a 47 growing demand for plastic products.⁸ 48

49 The first country that has recognized the importance and necessity to legislate and regulate the environmental issue of BPA was Canada.¹ Although, the European Commission 50 and the US government have banned the BPA use in baby bottles and packaging items for 51 52 children aged 0-3, BPA is still under discussion to be prioritized for regulation by the 53 European Union and the United States because of its possible adverse impact on human health and environment. Various declarations and statements have been issued [The Berlaymont 54 declaration, the Ramazzini statement, European Environmental Bureau, World Health 55 Organization, etc.] on this challenging topic, and the debate is still going on how to regulate 56 and reduce the BPA concentrations effectively in consumer products and aquatic systems in 57 order to protect humans and wildlife.¹ Recent studies confirmed the presence of BPA in 58 different human biological matrices such as urine⁹, breast milk¹⁰ and hair¹¹. 59

The main source of the BPA occurrence in the aquatic environment is municipal and 60 industrial wastewater. BPA is mostly found in water and food (including drinking water) 61 associated with the manufacture of various polycarbonate plastic products (packages or food 62 contact materials) or it is leached from the final products as a result of the conducted 63 hydrolysis of polymers at high temperature, or under acid/basic conditions.^{3,12,13} Despite its 64 rapid degradation, favoured by aerobic conditions, with a half-life of about 0.5-5 days in the 65 aquatic environment, the constant release into the surface water makes BPA a pseudo-66 persistent pollutant in the environment.¹⁴⁻¹⁶ 67

Although, European legislation set environmental quality standards for priority and certain other pollutants, BPA permitted levels, together with pharmaceuticals, personal care products, algal toxins and various endocrine disrupters have not been regulated by any legislative bodies yet because of the insufficient reliable data on their safety or risks.^{14,17}

In the Western Balkan Region, particularly in Serbia, only limited monitoring data on water status are available. A screening analysis of the wastewater discharged into the Danube along the Novi Sad municipality was carried out during the period 2011-2012, and more than 130 chemical contaminants were detected.¹⁸ Furthermore, non-target screening analysis of the Danube water samples in the area of Belgrade confirmed the presence of the wide range of organic contaminants.¹⁹

The primary goal of this study, conducted for the first time, was to examine the occurrence of BPA along the Danube River in the area of Novi Sad (Serbia). The obtained results were also used to evaluate the seasonal variations of this endocrine disrupter and to assess the potential ecological risks.

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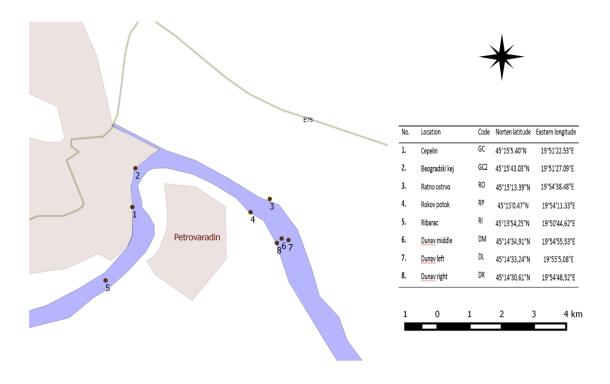
EXPERIMENTAL

83 Chemicals

The standard of bisphenol A, (2,3-dihydroxypropyl) glycid was obtained from Sigma Aldrich (Deisenhofen, Germany). Methanol, dichloromethane and ethyl acetate were of HPLC grade (LGC Promochem, Wesel, Germany). The silylation agent N,O-Bis(trimethylsilyl)trifluoroacetamide (BSTFA) was obtained from Fluka (Sigma Aldrich, Deisenhofen, Germany) and ultra pure water was used. The regenerated cellulose filters (0.45 µm) and cartridges were obtained from Agilent Technologies. All glassware was rinsed with methanol and dried in an oven at 100 °C for 1 h before use.

92 *Collection of water samples*

The surface water samples (1 L) were collected in dark amber glass bottles at eight representative locations along the Danube River (Fig. 1) of the Novi Sad municipality, Serbia, during autumn (November 2012), winter (March 2013), spring (May 2013) and summer (September 2013) under comparable meteorological conditions at all sites for each sample at the depth of 50 cm. The sampling locations, presented in Fig. 1, are in downstream order of the Danube River through the specified section.



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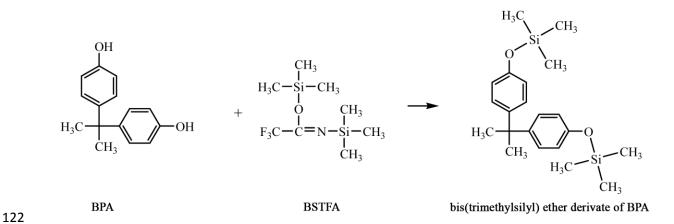
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Fig. 1. Map of the Danube River sampling sites.

101 The locations *GC*, *GC2*, *RO* and *RP* were positioned 100 meters downstream of 102 untreated sewage discharge. The sampling sites *DM*, *DL* and *DR* were collected few hundred 103 meters from the urban area to determine the eco-toxicological status of the river after the 104 effect of dilution and filtration by the river bank layers. The sampling site *RI* was included in 105 order to overview the influence of the incoming pollution to the Novi Sad section of the river. 106 All the samples were filtered through the regenerated cellulose 0.45 μm membrane filter 107 (Agilent Technologies, Germany) and extracted within 24 h in the laboratory.

108 *Sample preparation*

The sample preparation was done by using the previously developed method by 109 Ballesteros et al. (2006).²⁰ A solid-phase extraction procedure (SPE) with Agilent Bond Elut 110 Plexa (200 mg/6 mL) type of cartridges was used for the sample preconcentration and 111 removal of the existing interferents. SPE was carried out on an Agilent vacuum manifold for 112 12 columns connected to a water vacuum pump. The pH values of filtered water samples were 113 adjusted to pH 3 with 1M hydrochloric acid to increase the extraction efficiency.^{20,21} The 114 cartridges were activated with 5 mL mixture of dichloromethane and methanol (50:50, v/v) 115 followed with 5 mL of pure methanol and 15 mL of ultra pure water (pH 3). The final 116 extraction volume of each sample was 500 mL. The BPA was eluted with 5 ml of methanol. 117 After isolation, the eluate was gently evaporated to dryness under nitrogen stream at 40 °C 118 (5.0 purity) and the residue of all samples was reconstituted in 50 μ L of ethyl acetate. A 119 silvlation step was carried out using 50µL of BSTFA prior to the gas chromatography-mass 120 spectrometry (GC-MS) analysis (Fig. 2). 121



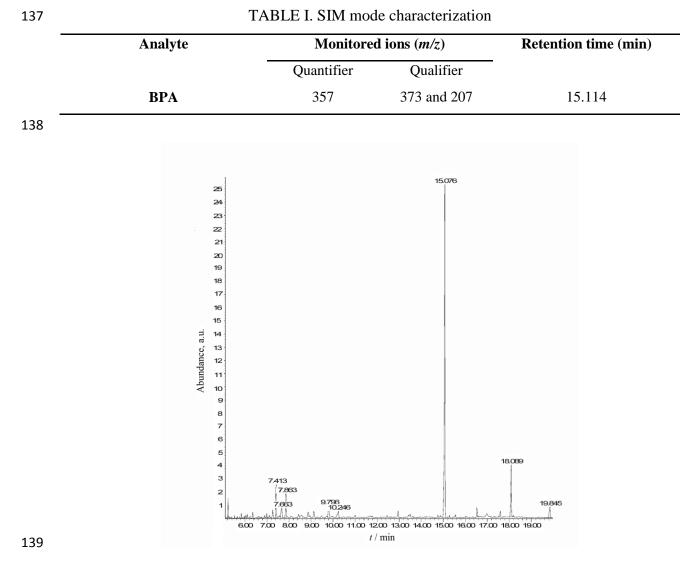
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124 *GC-MS analysis*

The analyses were done with an Agilent GC 7890A coupled with 5975C VL mass 125 detector using the method developed by Zafra et al. (2003).²² The system was used in SCAN 126 (complete spectrum) and in selected ion monitoring (SIM) modes. The separation was done 127 using a fused silica capillary column (30 m, 0.25 mm i.d., 0.25 µm film thickness; J&W 128 Scientific, Folsom, CA, USA). The MS quad temperature was set at 150 °C and the electron 129 impact (EI) ion source temperature of the MS was 230 °C. A 2 µL aliquot of the silvlated 130 samples were automatically injected using the splitless injection mode at 250 °C. The GC 131 oven temperature program was applied. The initial oven temperature was set at 50 °C and held 132

for 1 min. Then, the temperature was increased to 150 °C (held 1 min) and 270 °C via ramp of
25 and 10 °C/min respectively and maintained at 270 °C for 2 min. High purity helium (99.99
%) was used as carrier gas with a flow rate of 1.0916 mL/min. The characteristic ions listed in
Table I for SIM mode operation were used for the BPA quantification in the water samples.



140 Fig. 3. Representative GC-MS chromatogram of the Danube water sample (SIM mode).

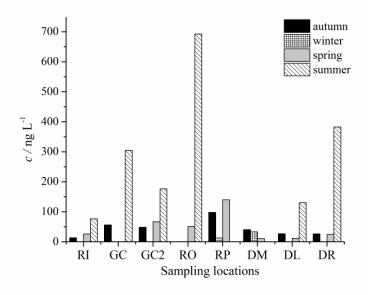
the water samples The BPA quantified with 141 in was external standard bis(trimethylsilyl)bisphenol A. The calibration curve was linear for the concentration range 142 35-100 ng/mL (y=40197x+10266, R²=0.997) and the limit of quantification (LOQ) was set at 143 S/N ratios \geq 10. All instrumental and procedural blanks were far below LOQ, 6 ng/L. The 144 identification of BPA was confirmed using Fiehn and NIST08 mass spectral libraries. The 145 characteristic chromatogram of the examined water samples containing BPA is presented in 146 147 Fig. 3.

RESULTS AND DISCUSSION

149 *Occurrence of BPA in surface water*

150 Seasonal occurrence of bisphenol A in the surface water at the chosen sites along the151 Danube of the Novi Sad section is shown in Fig. 4.

BPA was detected above the limit of quantification (LOQ), 6 ng/L, in 7 out of 8 collected samples in autumn. High frequency of BPA detection at concentration levels up to 98 ng/L confirmed the presence of the Danube surface water in this area. The maximum detected concentration was in the *RP* sample, which is near the numerous illegal settlements with septic tanks and waste disposal sites, unauthorized traffic and an old industrial complex.



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Fig. 4. Concentrations of BPA during the year.

In winter, BPA was only detected in two samples, *RP* and *DM*, at concentrations of 13 and 33 ng/L respectively, when the level of the Danube was increased as a result of the weather conditions. These concentrations are assumed to be due to the dilution effects. In winter the biodegradation by microorganisms should be slower. Two strains of *Pseudomonas* (*Pseudomonas sp.* and *Pseudomonas putida*) and *Stremtomyces sp.* strains were found to have BPA biodegradability capacity about 90% or more. The study conducted in China suggests that these bacteria are widely distributed in the rivers.¹⁴

However, high frequency of BPA detection was observed in the spring samples. Again, the maximum detected concentration of BPA was in the *RP* sample (140 ng/L), while the lowest concentration was in the *DM* sample (10.8 ng/L), the site which is far away from any

direct anthropological influence. The *DM*, *DL* and *DR* sites were chosen to determine if theNovi Sad area had an impact on the eco-toxicological status of the Danube.

The *RO* sampling site is located 100 m downstream of the discharge site of the sewage 171 collector for collecting the wastewater from the local big oil refinery complex and the rain and 172 urban run-off water from the industrial and underdeveloped suburban area. The possible 173 pollution of the main water supplying source of the Novi Sad drinking water waterworks by 174 the Oil Refinery and a Thermal Power Plant, situated only a few hundred meters upstream. 175 presents the major danger for the river.¹⁸ The detected BPA levels were above the LOQ in six 176 summer samples at significantly high concentrations, 77 ng/L in the RI sample and 693 ng/L 177 in the *RO* sample. 178

Although BPA has a short half-life in surface water due to the rapid degradation, this chemical was often detected in our study. The *RI* sampling point was included in order to overview the influence of the possible high incoming pollution as the Croatian biggest river port, Vukovar, is located 80 km upstream and presents the most important source of anthropological pollution of the Danube before Novi Sad. The BPA measured levels were above the LOQ (Table II) in autumn, spring and summer at this site.

185 The BPA concentrations in the surface water samples varied significantly depending on186 the chosen sites, possible sources and routes of pollution and the seasons (Table II).

Sample	Range (ng/L)	Mean (ng/L)	Median (ng/L)	95th percentile
RI	n.d77	29	20	77
GC	n.d304	90	28	304
GC2	n.d177	73	58	177
RO	n.d693	186	26	693
RP	n.d140	63	55	140
DM	n.d41	21	22	41
DL	n.d130	42	19	130
DR	n.d383	108	26	383

187	TABLE II. Seasona	d variation of BPA	concentrations in the	Danube surface v	vater samples

188 *n.d.:*<*LOQ*

The obtained BPA results in our study were compared with the available data on some European rivers. BPA was quantified in the Dutch surface waters at the levels up to 330 ng/L, but it was below the LOQ in 60-80% of water samples.²³ In the Seine River, BPA was found in the range of 11-154 ng/L²⁴, while in the Rhine surface water the concentration of BPA

ranged from 42 to 229 ng/L.²¹ However, in freshwater streams of Germany, the BPA was 193 detected in the concentrations up to 1927 ng/L, but only in 13% of the samples.²⁵ The 194 maximum detected BPA concentration in our study (693 ng/L) was significantly higher than 195 the obtained concentration in the recent publication analysing the Danube samples for this 196 endocrine disrupter (68 ng/L).²⁶ The relatively high BPA concentrations found in the 197 investigated Danube River stretch in Serbia can be explained by the fact that Novi Sad has 198 still no urban waste water treatment plant and that most of the industries around Novi Sad 199 municipality are without wastewater treatment facilities. 200

201 Risk evaluation

The presence of the BPA endocrine disrupter might have adverse effects on aquatic 202 organisms²⁷ and the consumption of BPA contaminated freshwater fish²³ is probably one of 203 the main routes of the contamination of humans.¹⁴ The environmental risk assessment is 204 evaluated by comparing the predicted environmental concentration (PEC) or by measuring the 205 environmental concentration (MEC) with the sensitivity of the aquatic ecosystem for BPA 206 (predicted no-effect concentration, PNEC). The ratio PEC/PNEC (or MEC/PNEC), known as 207 the risk quotient (RQ), should not exceed 1. Otherwise, a risk to the aquatic environment is 208 assumed.²⁸ 209

The PNEC values are based on eco-toxicity data from single species laboratory tests combined with the safety factors or even on the derived provisional PNEC.²⁹ Based on the literature data, the PNEC value of the BPA for fresh surface water varies from 64 μ g/L³⁰ to 1.5 μ g/L according to the European Union risk assessment report.³¹ However, the latest research of Wright-Walters et al. (2011)¹⁶ estimated an aquatic PNEC of 0.06 μ g/L for BPA using a non-parametric methodology.

In order to evaluate the BPA adverse effects on the Danube surface water, the 95th percentile of MEC was used together with PNEC to calculate the RQ for each sampling site using Wright-Walters and co-workers methodology.¹⁶

The calculated ratio ranged from 0.68 to 11.54 and the RQ was above 1 at seven locations out of eight (Fig. 5). Consequently, the frequency of exceedance was significantly high, 0.87. It is evident that a potential risk exists in this part of the Danube, especially near the RO sampling site.

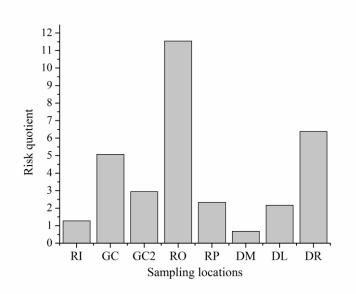




Fig. 5. Distribution of the eco-toxicological risk of BPA.

This risk assessment approach was based on an evaluation of the toxic potential of a single plasticizer to the aquatic environment without taking into account the possibility and effects of BPA acting additively with other EDs.

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CONCLUSIONS

This research is a pioneering study on the eco-toxicological impact of BPA as an 229 230 endocrine disrupter on the surface water in the Western Balkan Region. The obtained results suggest that the aquatic environment along the Danube River of the Novi Sad section, Serbia, 231 232 is not sufficiently protected from adverse effects to BPA exposure at the established water 233 concentrations of < 6 ng/L to 693 ng/L. The lack of the EDs legislation indicates that the 234 aquatic environment is not sufficiently protected from the detrimental effects of BPA toxicant. Constant and extensive monitoring of the high-risk sites, as well as the biological laboratory 235 experiments, are indispensable to get better insight into the potential toxicological risks of 236 BPA and understand the full scope of effects of the BPA concentration influence on aquatic 237 organisms along this part of the Danube. More reliable research on endocrine disruptive 238 substances such as BPA is necessary in the field of ecotoxicology (derivation of reliable 239 PNECs) and analytical chemistry (monitoring in the environment) in order to regulate these 240 241 xenoestrogens in the field of water policy.

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250	ИЗВОД		
251	СЕЗОНСКО ОДРЕЂИВАЊЕ ПРИСУСТВА БИСФЕНОЛА А У ПОВРШИНСКОЈ		
252	ВОДИ ДУНАВА		
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260	Сезонске промене концентрација бисфенола А у површинској води Дунава, на		
261	територији Новог Сада, анализиране су, по први пут на овом подручју, применом гасне		
262	хроматографије са масеном спектрометријом. И поред кратког полуживота у		
263	површинској води (0,5-5 дана), бисфенол А често је детектован дуж тока Дунава у		
264	релевантним концентрационим нивоима (до 693 ng/l). Регистроване вредности		
265	бисфенола А током четири годишња доба значајно се разликују и указују на већу		
266	контаминацију у летњем периоду.		
267	Процењен ризик по животну средину, изражен као однос 95. перцентила		
268	измерених концентрација бисфенола А у површинској води и концентрације за коју је		
269	предвиђено да нема биолошке ефекте, био је већи од јединице на чак седам локација		
270	узорковања.		
271	Добијени резултати у оквиру овог истраживања потврдили су загађење		
272	површинске воде Дунава бисфенолом А, дуж обала Новог Сада, и указују на постојање		
273	значајног ризика по људско здравље и животну средину у овом подручју.		
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